

Technical note

From : Govert de With (NRG)
To : Omri Lulav / / Kosta Kovler (NCAB)
Subject : Dose assessment from radon progeny in a concrete based room using CFD calculation
Datum : 25-1-2013
Referentie : 912559/13.118240 RE/GdW/VL

1 Radon simulation and boundary conditions

Geometry, material and ventilation conditions

Based on information provided by the customer the following conditions of the room have been agreed:

Geometry and material features of the room

Dimensions : length 3 m; width 3 m; height 2.7 m
Inner area : 50.4 m²
Volume : 24.3 m³
Window : 1.2 x 1.2 m² (1.4 m²)
Door : 2.0 x 0.8 m² (1.6 m²)
Area concrete : 47.4 m²
Concrete thickness : 0.2 m
Concrete density : 2300 kg·m⁻³
Furniture : Unfurnished room

Flow conditions and ²²²Rn background

Air exchange rate : 0.5 h⁻¹
²²²Rn background concentration : 10 Bq·m⁻³
²²²Rn background equilibrium factor : 0.4

²²²Rn source

According to exhalation measurements carried out by NRG^[1] the ²²²Rn exhalation rate is determined as 4.8·10⁻⁶ and 6.8·10⁻⁶ Bq·kg⁻¹·s⁻¹ for the concrete samples with and without fly ash, respectively. In building materials the diffusion of ²²²Rn is often described as a one-dimensional process, i.e. directed to the living space and its opposite side. The ²²²Rn source for the considered room follows from the multiplication of the exhalation rates by the total area in the room taken by concrete and the surface density, the product of the thickness of the building parts and their density. In formula:

$$S = E_m A \rho L \quad (1)$$

In this equation S represents the ²²²Rn source in the room (Bq·s⁻¹); E_m the ²²²Rn exhalation rate (Bq·kg⁻¹·s⁻¹); A the area in the room taken by concrete (m²); ρ the density of concrete (kg·m⁻³); and L the half-thickness of the construction parts (m). In this way the ²²²Rn source in the room is calculated at 7.4·10⁻² Bq·s⁻¹ for the room constructed with concrete to which no fly ash is added and 5.2·10⁻² Bq·s⁻¹ for the concrete with fly ash.

These values, however, have to be considered as a maximum. The exhalation from the test specimen takes place from all sides instead of one-dimensional as in a practical situation. This may lead to an overestimation of the ^{222}Rn source in the room under investigation. Berkvens et al.^[2] have introduced a formalism to calculate the one-dimensional exhalation rate from the experimentally determined results. Therefore, they defined an equivalent sample geometry with a reduced thickness to compensate for the ^{222}Rn diffusion to the planes other than the front and backside. With the x-axis directed towards the room, the shortest mean distance is given by:

$$\frac{L'_x}{2} = \frac{L_x}{2} - \frac{1}{6} \left(\frac{L_x^2}{L_y} + \frac{L_x^2}{L_z} - \frac{1}{2} \frac{L_x^3}{L_y L_z} \right) \quad (2)$$

In this expression L_x , L_y and L_z represent half of the sample thickness of the test specimen in the x-, y- and z-direction. As the dimensions of the present test samples are $0.10 \times 0.10 \times 0.20 \text{ m}^3$, L'_x is calculated at 0.33 m. Subsequently the exhalation rate from a wall in practice is calculated from the measured exhalation rate according to:

$$\frac{E_m^{(1)}}{E_m} = \frac{L'_x \tanh(L_x/l)}{L_x \tanh(L'_x/l)} \quad (3)$$

$E_m^{(1)}$ in this equation is the one-dimensional exhalation rate; E_m the experimentally determined exhalation rate; and l the diffusion length of ^{222}Rn in concrete. As reported by Kovler et al.^[3] the literature data on the diffusion length in concrete ranges from 0.06 m for heavy concretes to 0.30 m for light-weight species. If for the present concretes a diffusion coefficient of 0.10 m is assumed, a correction factor of 0.79 can be calculated from equation (3). With that the ^{222}Rn source in the considered room comes to $5.9 \cdot 10^{-2} \text{ Bq} \cdot \text{s}^{-1}$ (no fly ash added) and $4.1 \cdot 10^{-2} \text{ Bq} \cdot \text{s}^{-1}$ (with fly ash), respectively. Since these values are supposed to be more realistic, these are applied in the model calculations.

Radon and radon progeny modelling

The modelling of the ^{222}Rn concentration and the concentrations of the short-lived ^{222}Rn progeny nuclides ^{218}Po , ^{214}Pb and ^{214}Bi follows the method as described De With and De Jong^[4]. In this work, a Computational Fluid Dynamics (CFD) model is used to simulate the concentration of thoron and thoron progeny products in a typical Dutch living room. The dispersion is computed using the fundamental flow equations for gas and aerosols, which enables detailed simulation of the three-dimensional flow structures from ventilation and buoyancy. Extra algorithms are developed and coupled with the CFD model to take account of all relevant physical processes. These include the formation and attachment of the progeny products to aerosol particles as well as their dispersion and deposition.

The computed concentrations for radon and its progeny are as follows:

Table 1 Concentrations radon and radon progeny.

	$^{222}\text{Rn} (\text{Bq} \cdot \text{m}^{-3})$	$^{218}\text{Po} (\text{Bq} \cdot \text{m}^{-3})$	$^{214}\text{Pb} (\text{Bq} \cdot \text{m}^{-3})$	$^{214}\text{Bi} (\text{Bq} \cdot \text{m}^{-3})$
Concrete without fly ash	27.3	23.3	9.9	5.1
Concrete with fly ash	22.2	19.0	8.2	4.3

These concentration levels also include the radon background concentration of $10 \text{ Bq} \cdot \text{m}^{-3}$ and its background progeny.

2 Dose assessment

Calculation of internal dose

The effective dose to residents due to inhalation of the short-lived ^{222}Rn progeny is based on the so-called equilibrium equivalent ^{222}Rn concentration. This concentration is calculated according to the following expression:

$$C_{eq}^{222} = 0.105 C_1 + 0.515 C_2 + 0.380 C_3 \quad (4)$$

in which C_1 , C_2 and C_3 are the activity concentrations in the indoor environment of ^{218}Po , ^{214}Pb and ^{214}Bi ($\text{Bq}\cdot\text{m}^{-3}$), respectively, as found by the model computations. The constants in the equation are the relative contributions of each decay product to the total potential alpha energy from the decay of a unit ^{222}Rn gas. To convert the equilibrium equivalent concentration into an effective dose rate, a conversion factor of $9 \text{ nSv}\cdot\text{h}^{-1}$ per $\text{Bq}\cdot\text{m}^{-3}$ is applied^[5]. The time spent indoors is taken as 7000 h per year (80% of the total time).

Based on the findings reported in Table 1, the above conversion factor and the hours spent indoors the internal dose is as follows:

Table 2 Annual internal dose from radon exposure.

	Internal dose ($\text{mSv}\cdot\text{a}^{-1}$)
Concrete without fly ash	0.60
Concrete with fly ash	0.50

Calculation of external dose

The external dose due to gamma radiation from the building materials in the room is calculated according to the method described by De Jong et al.^[6,7]. The standard geometry as defined by Koblinger^[8] is taken as starting point for our calculation model. This geometry has dimensions of $5 \times 4 \text{ m}^2$ and 2.8 m in height, with each construction part (i.e. floor, walls and ceiling) made of 20 cm thick concrete and no doors or windows. Correction factors have been deduced for alternative situations. The absorbed dose rate in air (unit: $\text{Gy}\cdot\text{h}^{-1}$) in a particular room is then calculated according to:

$$\dot{D}_{air} = \left\{ \sum_{i=1}^6 [F_{dose} \cdot F_1 \cdot F_2 \cdot F_3 \cdots F_n]_i \right\} F_{zoning} \cdot F_{adjac} \quad (5)$$

in which i is the index for a construction part, F_1 to F_n are the correction factors for construction part i , F_{zoning} is a correction factor which takes internal zoning of the construction into account, and F_{adjac} is the contribution from adjacent floors and dwellings. F_{dose} is the so-called dose factor, defined as:

$$F_{dose,i} = k_1 a_{1,i} + k_2 a_{2,i} + k_3 a_{3,i} \quad (6)$$

In this equation k_1 , k_2 and k_3 represent the specific absorbed dose rates, defined as the absorbed dose rate in air due to an activity concentration of $1 \text{ Bq}\cdot\text{kg}^{-1}$ of each of the primordial radionuclides in equilibrium with its decay products ($\text{Gy}\cdot\text{h}^{-1}$ per $\text{Bq}\cdot\text{kg}^{-1}$); and $a_{1,i}$ to $a_{3,i}$ the activity concentration of ^{226}Ra , ^{232}Th and ^{40}K of construction part i ($\text{Bq}\cdot\text{kg}^{-1}$), respectively. The values of the specific

absorbed dose rate depend amongst others on the thickness, density and dimensions of the various construction parts.

Several researchers have determined the specific absorbed dose rate for the standard Koblinger-construction, using various codes. Averaged over all data the specific absorbed dose rates are^[9]:

²²⁶ Ra	0.90	nGy·h ⁻¹ per Bq·kg ⁻¹
²³² Th	1.10	nGy·h ⁻¹ per Bq·kg ⁻¹
⁴⁰ K	0.08	nGy·h ⁻¹ per Bq·kg ⁻¹

For the calculation of the annual indoor effective dose to the residents due to external radiation the absorbed dose as found from equation (5) is multiplied by a conversion factor of 0.7 Sv·Gy⁻¹ and the annual number of hours spent indoors, taken as 7000. The required activities for each of the three radionuclides are based on measurements performed by NRG^[1]. The mean activity concentrations for both types of concrete are as follows:

Table 3 Activity concentrations.

	²²⁶ Ra (Bq·kg ⁻¹)	²³² Th (Bq·kg ⁻¹)	⁴⁰ K (Bq·kg ⁻¹)
Concrete without fly ash	29.4	8.6	56.1
Concrete with fly ash	37.9	18.1	60.9

Based on those experimental results the annual external dose for is as follows:

Table 4 Annual external dose from primordial radionuclides.

	Internal dose (mSv·a ⁻¹)
Concrete without fly ash	0.20
Concrete with fly ash	0.29

Calculation of total dose

Based on the calculated internal and external dose as presented in Table 2 and Table 4 the total annual dose per year is:

Table 5 Annual total dose from primordial radionuclides.

	Total dose (mSv·a ⁻¹)
Concrete without fly ash	0.80
Concrete with fly ash	0.78

3 References

- [1] De Jong, P. (2010)
Analysis of radon exhalation rate and activity concentrations. Letter K5098/10.104560 RE/PdJ/VL, Nuclear Research and consultancy Group.
- [2] Berkvens, P., Kerkhove, E. and Vanmarcke, H. (1988)
Three-dimensional treatment of steady-state ^{222}Rn diffusion in building materials: introducing a practical modified one-dimensional approach. *Health Phys.* 55:793-799.
- [3] Kovler, K., Perevalov, A., Steiner, V. and Rabkin, E. (2004)
Determination of the radon diffusion length in building materials using electrets and activated carbon. *Health Phys.* 86:505-516.
- [4] De With G. and De Jong P. (2011).
CFD modelling of thoron and thoron progeny in the indoor environment. *Rad. Prot. Dosim.*, 145 (2-3): 127-132.
- [5] UNSCEAR (2000).
Sources and effects of ionizing radiation. UNSCEAR 2000 report to the General Assembly, with scientific annexes. Volume 1: Sources. United Nations Scientific Committee on the Effects of Atomic Radiations. United Nations, New York.
- [6] De Jong, P. and Van Dijk, W. (2008)
Modeling gamma radiation in dwellings due to building materials. *Health Phys.* 94: 33-42.
- [7] De Jong, P. and Van Dijk, J.W.E. (2008)
Calculation of the indoor gamma dose rate distribution due to building materials in the Netherlands. *Radiat. Prot. Dosim.* 132:381-389.
- [8] Koblinger, L. (1978)
Calculation of exposure rates from gamma sources in walls of dwelling rooms. *Health Phys.* 34: 459-463.
- [9] De Jong, P. (2010)
PhD Thesis Exposure to natural radioactivity in the Netherlands: the impact of building materials, Arnhem, The Netherlands.